JEE Journal of Ecological Engineering

Volume 20, Issue 5, May 2019, pages 203–208 https://doi.org/10.12911/22998993/105474

Advanced Oxidation Treatment of Composting Leachate of Food Solid Waste by Ozone-Hydrogen Peroxide

Maciej Gliniak^{1*}, Anna Lis¹, Daria Polek², Marta Wołosiewicz-Głąb²

¹ University of Agriculture in Krakow, al. Mickiewicza 21, 31-120 Krakow, Poland

² AGH University of Science and Technology, al. Mickiewicza 30, 30-059 Krakow, Poland

* Corresponding author's e-mail: maciej.gliniak@urk.edu.pl

ABSTRACT

The research was conducted to investigate the efficiency and possibilities of advanced oxidation process based on ozone-hydrogen peroxide. The process was used as a post-treatment step of composting leachate utilisation. The leachate samples were collected from a typical composting plant with the aerobic biological treatment system. The samples were conditioned in a "ozone reactor" without dilution. The effectiveness of the treatment process was measured by pH values (4.0-7.0), H_2O_2 concentrations $(0.5-4.0 \text{ g}\cdot\text{dm}^{-3})$, ozone doses $(0.5-1.5 \text{ g}\cdot\text{h}^{-1})$ and reaction times (0-10 min). The highest removal efficiencies achieved were 85% and 92% for chemical oxygen demand (COD) and biochemical oxygen demand (BOD₅), respectively. The optimum process parameters occurred at pH 5.0, 2.0 g·dm⁻³ concentration of H_2O_2 , and 0.75 g·h⁻¹ of ozone dose. The optimal reaction time was 3 min. The O_3/H_2O_2 advanced oxidation process was found to oxidize COD and BOD₅ of the composting leachate. The oxidation reaction can be used as a feasible technique for composting leachate treatment.

Keywords: BOD₅, COD, oxidation, compost, leachate

INTRODUCTION

Food wastes (FW) are the main component of municipal solid waste (MSW). The estimates provided by Food and Agricultural Organization (2011), Carlson et al. (2012) and Zaman (2013) indicate increasing global food consumption and an over 40% increase in FW production by 2025. While analyzing the issues of the MSW economy, attention should be paid to the increase in the recycling of biodegradable waste. This process is related to the progressive reduction of storing in landfills waste that can be re-used (Wiliams, 2005).

One of the most popular FW recycling methods is composting. It belongs to the key technologies commonly used worldwide (Pitchel, 2010). Under temperate climate conditions, the composting process is divided into two stages, which in total should last for at least 8 weeks. The long duration time of the process depends on the choice of the composting method (active system

with oxidation of the material or passive without oxidation) and the processed material type. On the basis of the research conducted in various countries, it was found that the first phase of the composting process should last approximately 2 weeks. Then, the material is directed to stabilization, which lasts 6-8 weeks. Currently, many studies in the field of optimization of the kinetics of the composting process in its first phase and the elimination of accompanying odorous substances are conducted globally (Lebrero et al., 2011; Yuan et al., 2015; Fernández et al., 2016; Jinyi et al., 2016; Siles et al., 2016; Yongjiang et al., 2016). In addition, in the literature on the subject there are studies on the additional hygienization of the stabilized material and the possibility of using gases generated during the process. One of the main problems associated with composting is the generation of the leachate containing material that has been extracted, suspended or dissolved in waste (Benlboukht et al., 2016; Junya et al., 2016; Lian et al., 2016; Mukesh et al., 2016).

advanced oxidation processes (AOP)are very popular in current technologies of neutralization of fluid process residues. Most commonly, they are used to eliminate color, reduce the load of organic pollutants or reduce the toxicity of water and wastewater (Jia et al., 2011; Cortez et al., 2011). The principle of AOP is based on the release of hydroxyl radicals, which accelerate the degradation of organic compounds in the aquatic environment.

The most popular AOP is the O_2/H_2O_2 system, which is used as a chemical process for purification and pre-treatment of wastewater (Maniero et al. 2008; Rosal et al., 2009; Liu et al., 2009; Quiang et al.; 2010). Tizaoui et al. (2007) obtained during the ozonization of the leachate the COD removal rate of 27% and the color reduction by 85%. Rivas et al. (2003) and Hagman et al. (2008) received at least a 20% reduction in COD using only hydrogen peroxide (H₂O₂). Using ozone along with the H₂O₂, a COD reduction of 20-50% can be achieved (Tizaoui et al., 2007; Hagman et al., 2008; Goi et al., 2009). The presented results refer to the application of the AOP system for wastewater. The leachates from the composting process are a more specific waste, the properties of which depend mainly on the oxygenation and moisture of the processed material.

The purpose of this article is to present the possibility of using the O_3/H_2O_2 oxidation system in the process of preliminary preparation of the leachate from the FW composting process for neutralization in a professional wastewater treatment plant. The effectiveness of the process was assessed by the BOD₅, COD dynamics of change and BOD₅/COD ratio.

MATERIAL AND METHODS

Leachate samples

The leachate samples were prepared during the FW composting process in the BKB 100 isothermal bioreactor for the study of organic material decomposition processes. A FW mixture consisting of vegetables (50%), fruit (25%), expired food (15%) and sawdust (10%) with a total fresh weight of 50 kg was used for the tests. The composition of the mixture was selected to obtain appropriate structural properties of compost, while ensuring proper air flow in the bioreactor, biomass moisture during the process (54–68%) and a C:N ratio above 20. In order to achieve a

Experimental procedure

The advanced leachate oxidation process was carried out in a flow glass reactor. The reactor is equipped with a stirrer, by means of which it is possible to aerate the mixture with the ozone coming from the discharge generator. Through the system of independent infusions, the leachates and the hydrogen peroxide solution are infused into the reaction chamber. The AOP procedure proceeded according to the following scheme:

- 1. Filling the reactor with 1 dm³ of the leachate, which was mixed applying air with a capacity of 5.0 dm³·min⁻¹ over a period of 5 minutes.
- 2. The AOP process involved:
- a) Addition of H_2O_2 (1 g·dm⁻³) and ozone (1 g·h⁻¹), reaction time 10 minutes, regulation of the reaction using 1N sodium hydroxide in the range of 4÷7 (optimization of the reaction).
- b) Differentiated H₂O₂ concentration at the optimal pH in the range of 0.5÷4 g·dm⁻³, ozone dose 1 g·h⁻¹, reaction time − 10 minutes.
- c) Modification of ozone concentration in the range of $0.5 \div 1.5 \text{ g} \cdot \text{h}^{-1}$, reaction time 10 minutes, optimal pH and H_2O_2 concentration.
- 3. For each step presented in section 2, pH, COD, BOD₅, BOD₅/COD analyses were performed.

Materials and analysis

Hydrogen peroxide (30%, w/w) and sodium hydroxide solution (1 N) were of analytical reagent grade (Merck, Germany). Ozone was produced in the O3PRO30,7VW generator, equipped with a corona electrode system. The pH was measured by an Elmetron CP-511 conductometer with an EPS-1 glass electrode for the measurement of pH in the aqueous solution. COD was measured using a standard miniaturized method in airtight tubes, in accordance with the PN-ISO 15705:2005 standard. For the BOD₅ measurement, the respirometric method described in DIN EN 1899 H55 was used, which uses OxiTop Control IS (WTW, Germany) for measurements. All experiments were carried out in 5 repetitions. During testing, the temperature was maintained at 22±1°C and humidity at 55±5%.

The examined leachate samples were taken from the process of composting biodegradable waste in container technology. The basic leachate parameters were 36.9 g·dm⁻³ (COD), 19.7 g·dm⁻³ (BOD₅) at pH 7.32 and electrical conductivity 12.73 mS·cm⁻¹. The sample of the leachate was characterized by a black color and a sharp, irritating odor. The analyzed leachate samples have distinctively high level of oxygen demand, whereas the BOD₅/COD ratio was 0.53, which indicates a high potential of biodegradability of organic matter.

The values of COD and BOD₅ at the pH of 5.0 reached maximum, which amounted to 26% and 31%, respectively (Fig. 1). The efficiency of oxygen demand reduction fell above the pH value of 9.5 and remained stable. The Pearson correlation test showed a direct linear relationship between the pH values and COD removal efficiency (P<0.05, r=0.817) and BOD₅ (P<0.05, r=0.836).

The effect of various concentrations of H_2O_2 was examined in eight variants of concentration $0.5 \div 4 \text{ g} \cdot \text{dm}^{-3}$ with the ozonization efficiency of 1 $\text{g} \cdot \text{h}^{-1}$ (Fig. 2). The maximum degradation of COD and BOD₅ was 42% and 49%, respectively, at a concentration of 3.0 $\text{g} \cdot \text{dm}^{-3} H_2O_2$. A further increase in the H_2O_2 concentration caused slower degradation of both COD and BOD₅. The Pearson correlation test showed a direct linear relationship between the H_2O_2 concentration and BOD₅ removal efficiency (P < 0.05, r = 0.728). At the same time, this test showed an inverse relationship between H_2O_2 concentration and COD removal, which was not significant (P < 0.05, r = -0.408).

The effect of the ozone dose on the removal of COD and BOD, in the presence of an optimized

dose of H_2O_2 and pH (5.0) caused a linear increase in the degradation of oxygen demand and constant reaction time (Fig. 3). The highest efficiency of COD removal (36%) was achieved with 0.75 g·dm⁻³ ozone dose. Removal of BOD₅ achieved the highest reduction at an analogous dose of ozone and amounted to 49%. The above-mentioned results were achieved during a 10-minute reaction time.

In order to examine the effect of reaction time on the oxygen demand reduction efficiency, an experiment was carried out at variable reaction times of 0÷10 minutes in the previously optimized doses of H₂O₂ and O₃ (Fig. 4). The results showed that the increase in the reaction time linearly increases the efficiency of oxygen demand reduction. The maximum COD and BOD, reduction values were 63% and 69%, respectively, at 4 minutes. In the reaction time range of $6\div10$ minutes, the oxygen demand reduction efficiencies did not change significantly. The results of Pearson correlation test showed a direct linear relationship between the reaction time and the reduction efficiency of BOD_c (P < 0.05, r = 0.992) and COD (P < 0.05, r = 0.998).

The BOD₅/COD ratio during the ozonization changes in time analogous to the variability of the reduced indicators. For the untreated samples, it reached 0.53. After taking into account the optimal process conditions, the BOD₅/COD ratio amounted to 0.38.

DISCUSSION

In the carried out study, stable results were obtained at lower pH values than described in 20–21. Figure 1 shows that a more stable



Fig. 1. Effect of the pH on COD and BOD, removal (H₂O, $[1.0 \text{ g} \cdot \text{dm}^{-3}]$, O, $[1.0 \text{ g} \cdot \text{h}^{-1}]$ and reaction time [10 min])

reduction in oxygen demand was achieved by leachate than in Cortez et al. (2011). The studies showed that the effect of using H_2O_2 positively influenced the efficiency of removing pollutants while using concentrations in the range of 1.5–2.5 g·dm⁻³. Cortez et al. (2011) noted the positive effect of using the O_3/H_2O_2 system in relation to the ozonation alone. In his studies, he achieved a reduction of COD from 27% to 72% depending on hydrogen peroxide concentration.

Studies by Tizaoui et al. (2007) and Rivas et al. (2003) stated that the application of excessive doses of H_2O_2 adversely affects the process of wastewater treatment. This has also been confirmed in the carried out studies as illustrated in Figure 2. Tizaouni et al. (2007) received optimal process parameters at the hydrogen peroxide concentration 2 g·dm⁻³. With further increase in concentration (up to 6 g·dm⁻³) the process efficiency decreased by 6.5%.

Ozone may participate in the generation of hydroxyl radicals. While increasing the concentration of O_3 in the studied process, the COD and BOD₅ removal rate improved (Fig. 3). Hydroxyl radicals that have not participated in wastewater treatment, together with ozone, increase the degradability of pollutants, confirming studies 24. In the study by Maniero et al. (2008) a 70% decrease

in the COD value was obtained after the application of ozone at a dose of 9 g·dm⁻³. In analogous studies by Cortez et al. (2011) COD removal at 25% was achieved with the concentration of O_3 in the range of 63–112 g·dm⁻³.

Reaction time is a key indicator of the wastewater treatment process. The conducted research has shown that the use of AOP effectively reduces the time needed to remove contaminants. At the same time, the positive, the synergistic effect of using oxidants on wastewater treatment, described by 8–20, was confirmed. The achieved results of reducing the reaction time (Fig. 4), gave better results than those described in papers Lian et al. (2016) and Mukesh et al. (2016).

The presented research results show that the proper use of the O_3/H_2O_2 oxidation system increases the biodegradability of wastewater. Similar results were obtained by Goi et al. (2009) during the leachate treatment using the Fenton reaction. In addition, during the process of contaminants oxidation, changes in the molecular structure of contaminants were achieved, as shown by the BOD₅/COD indicator. By reducing the oxygen demand, a transformation of compounds that are difficult to biodegrade into more degradable forms has been achieved, which was also described by Jinyi et al. (2016) and Rivas at al. (2003).



Fig. 2. Effect of the H₂O₂ on COD and BOD₅ removal (pH [5.0], O₃ [1.0 g·h⁻¹] and reaction time [10 min])



Fig. 3. Effect of the O₃ on COD and BOD₅ removal (pH [5.0], H₂O₇ [3.0 g · dm⁻³] and reaction time [10 min])



Fig. 4. Effect of the reaction time on COD and BOD, removal (pH [5.0], H_2O_2 [1.0 g·dm⁻³] and O_3 [0.75 g·h⁻¹])

CONCLUSION

The results of the carried out tests clearly show that the combination of H_2O_2 and O_3 in one oxidation system is an effective way to reduce the oxygen demand during leachate treatment. The optimum process conditions occur at the pH of 5.0 and the addition of oxidants at the level of 0.75 g·dm⁻³ each. The optimal and reasonable reaction time is 4 minutes. The test results indicate the beneficial effect of AOP on process leachates, reducing the BOD₅/COD ratio by 30%, thus – lowering the oxygen demand during the treatment. The use of this reaction, due to the high biodegradability of leachates from the composting process, can be a preparatory process before transferring them to a wastewater treatment plant.

Acknowledgements

Publication was financed by the Ministry of Science and Higher Education of the Republic of Poland – statutory activity no. BM-4639 and DS3600/WIPiE/2018.

Publication supported by the Polish Ministry of Science and Higher Education as a part of the program of activities disseminating science from the project "Organization of the First International Science Conference – Ecological and Environmental Engineering", 26–29 June 2018, Kraków.

REFERENCES

 Benlboukht F., Lemee L., Amir S., Ambles A., Hafidi M. 2016. Biotransformation of organic matter during composting of solid wastes from traditional tanneries by thermochemolysis coupled with gas chromatography and mass spectrometry. Ecological Engineering, 90, 87-95.

- Carlsson M., Lagerkvist A., Morgan-Sagastume F. 2012. The effects of substrate pretreatment on anaerobic digestion: a review. Waste Management, 32, 1634–1650.
- Cortez S., Teixeira P., Oliveira R., Mota M. 2011. Evaluation of Fenton and ozone-based advanced oxidation processes as mature landfill leachate pretreatments. Journal of Environmental Management, 92, 749–755.
- DIN EN 1899 H55. 1998. Water quality Determination of biochemical oxygen demand after n days (BODn) – Part 2: Method for undiluted samples. International Organization for Standardization.
- Fernández C., Mateu C., Moral R., Sole-Mauri F., 2016. A predictor model for the composting process on an industrial scale based on Markov processes. Environmental Modelling & Software, 79, 156–166.
- Food and Agricultural Organization. 2011. Global Food Losses and Food Waste – Study Conducted for the International Congress. Swedish Institute for Food and Biotechnology.
- Goi A., Veressinina Y., Trapido M. 2009. Combination of ozonation and the Fenton processes for landfill leachate treatment: Evaluation of treatment efficiency. Ozone: Science & Engineering, 31, 28–36.
- Hagman M., Heander E., Jansen J.L. 2008. Advanced oxidation of refractory organics in leachate

 Potential methods and evaluation of biodegradability of the remaining substrate. Environmental Technology, 29, 941–946.
- Jia C., Wang Y., Zhang C., Qin Q. 2011. UV-TiO2 photocatalytic degradation of landfill leachate. Water, Air, & Soil Pollution, 217, 375–385.
- Jinyi G., Guangqun H., Jing H., Jianfei Z., Lujia H. 2016. Manure–wheat straw composting: A new approach that considers surface convection. International Journal of Heat and Mass Transfer, 97, 735–741.
- Junya Z., Meixue Ch., Qianwen S., Juan T., Chao J., Xueting L., Yuxiu Z., Yuansong W. 2016. Im-

pacts of addition of natural zeolite or a nitrification inhibitor on antibiotic resistance genes during sludge composting. Water Research, 91, 339–349.

- Lebrero R., Rodríguez E., García-Encina P., A., Muñoz R. 2011. A comparative assessment of biofiltration and activated sludge diffusion for odour abatement. Journal of Hazardous Materials, 190(1–3), 622–630.
- Lian Y., Shihua Z., Zhigiang Ch., Qinxue W., Yao W. 2016. Maturity and security assessment of pilot-scale aerobic co-composting of penicillin fermentation dregs (PFDs) with sewage sludge. Bioresource Technology, 204, 185–191.
- Liu H., Liang M.Y., Liu C.S., Gao Y.X., Zhou J.M. 2009. Catalytic degradation of phenol in sonolysis by coal ash and O3/H2O2. Chemical Engineering Journal, 153, 131–137.
- Maniero M.G., Bila D.M., Dezotti M. 2008. Degradation and estrogenic activity removal of 17betaestradiol and 17alpha-ethinylestradiol by ozonation and O3/H2O2. Science of the Total Environment, 407, 105–115.
- 16. Mukesh K.A., Akhilesh K.P., Pushpendra S.B., Wong W.C., Li R., Zengqiang Z. 2016. Co-composting of gelatin industry sludge combined with organic fraction of municipal solid waste and poultry waste employing zeolite mixed with enriched nitrifying bacterial consortium. Bioresource Technology, 213, 181–189.
- Pichtel J. 2010. Waste Management Practices: Municipal, Hazardous, and Industrial, second edition. Taylor and Francis, New York.
- PN-ISO 15705. 2002. Water quality Determination of the chemical oxygen demand index (ST-COD) – Small-scale sealed-tube method. International Organization for Standardization.
- 19. Qiang Z., Liu C., Dong B., Zhang Y. 2010. Degradation mechanism of alachlor during direct ozo-

nation and O3/H2O2 advanced oxidation process. Chemosphere, 78, 517–526.

- Rivas F.J., Beltrán F., Gimeno O., Acedo B., Carvalho F. 2003. Stabilized leachates: Ozone-activated carbon treatment and kinetics. Water Research, 37, 4823–4834.
- Rosal R., Rodríguez A., Perdigón-Melón J.A., Petre A., García-Calvo E., 2009. Oxidation of dissolved organic matter in the effluent of a sewage treatment plant using ozone combined with hydrogen peroxide (O3/H2O2). Chemical Engineering Journal, 149, 311–318.
- 22. Siles J.A., Vargas F., Gutiérrez M.C., Chica A.F., Martín M.A. 2016. Integral valorisation of waste orange peel using combustion, biomethanisation and co-composting technologies. Bioresource Technology, 211, 173–182.
- Tizaoui C., Bouselmi L., Mansouri L., Ghrabi A. 2007. Landfill leachate treatment with ozone and ozone/hydrogen peroxide systems. Journal of Hazardous Materials, 140, 316–324.
- 24. Williams P.T. 2005. Waste Treatment and Disposal. John Wiley and Sons, second edition, Great Britain.
- 25. Yongjiang W., Li P., Xinyu L., Yuansheng W., Kexun Z., Fei L. 2016. Using thermal balance model to determine optimal reactor volume and insulation material needed in a laboratory-scale composting reactor. Bioresource Technology, 206, 164–172.
- 26. Yuan J., Yang Q., Zhang Z., Li G., Luo W., Zhang D. 2015. Use of additive and pretreatment to control odors in municipal kitchen waste during aerobic composting. Journal of Environmental Sciences, 37, 83–90.
- Zaman A.U. 2013. Identification of waste management development drivers and potential emerging waste treatment technologies. International Journal of Environmental Science and Technology, 10(3), 455–464.